



# Anthropogenic fibers and microplastics in the pelagic gooseneck barnacle *Lepas (Lepas) anatifera* in Capo Milazzo Marine Protected Area (Tyrrhenian Sea): A first characterization

Gianfranco Scotti<sup>a</sup>, Michela D'Alessandro<sup>b,\*</sup>, Valentina Esposito<sup>b</sup>, Pietro Vivona<sup>a</sup>, Cristina Panti<sup>c</sup>

<sup>a</sup> Institute for Environmental Protection and Research, ISPRA via dei Mille 46, 98057 Milazzo, ME, Italy

<sup>b</sup> National Institute of Oceanography and Applied Geophysics – OGS, via Auguste Piccard 54, 34151 Trieste, Italy

<sup>c</sup> Department of Physical Sciences, Earth and Environment, University of Siena, Via P. A. Mattioli, 4, 53100 Siena, Italy

## ARTICLE INFO

### Keywords:

Fibers  
Microplastics  
Marine pollution  
*Lepas (Lepas) anatifera*  
Mediterranean Marine Protected Area

## ABSTRACT

This study provides an assessment of the efficiency of the biofouler *Lepas (Lepas) anatifera* Linnaeus, 1758 in capturing microplastics and microfiber particles floating in the water column. In this context, we collected pelagic gooseneck barnacles at fixed moorings in the Capo Milazzo Marine Protected Area (MPA). Fibers and fragments were found in the digestive tract of 30% of the 120 specimens collected. The ingested debris were mainly fibers (85.9%) of synthetic (30.6%) and natural (11.7%) origin, with length ranging between 1 and 2 mm (33.3%) and transparent (47.2%). The highest concentration of fibers was recorded in barnacles collected in the western sector of the MPA that is more affected by the input of organic matter coming from several urban wastewaters. No correlation was found between the presence of artificial polymers in *L. (Lepas) anatifera* and the size of individuals. The great diversity of fibers and plastic fragments by typology, size, shape and color, as well as the large number of bivalve and ostracod shells found in the digestive tract of the samples, confirm the opportunistic diet of these organisms, which can indiscriminately ingest any type of prey and cannot distinguish between microplastics and food. The results obtained, as well as the wide distribution and abundance, and ease of sampling of these barnacle species in macrofouling suggest that including this species in monitoring programs could be a cost-effective and easy method for assessing the presence of microplastics and microfibers in coastal marine waters to monitor the ecological status of pelagic and coastal ecosystems, including MPAs.

## 1. Introduction

Microplastics include a wide range of particles, <5 mm in size, that differ in shape, color and polymer. They can have different origin and sources and may result from multiple production processes and anthropogenic activities (Lima et al., 2021; Sharma et al., 2021). Many of these particles contain additives, which have been added to the original natural or synthetic material, to improve their performance and prolong their life (Hahladakis et al., 2018). Stabilizing additives, plasticizers, pigments, flame retardants, antioxidants and antimicrobials can leach from the plastic material and spread into the environment. For these reasons, microplastics, with any associated additive, are considered as environmental pollutants (Gunaalan et al., 2020; Teuten et al., 2009) and require a comprehensive risk assessment (Burns and Boxall,

2018; Gunaalan et al., 2020). The two most common forms of microplastics in the aquatic environment are small fragments (particles) and fibers (microfibers) (Barrows et al., 2018). Studies on the presence of microplastics in natural environments evidenced that fibers (>5 mm) (synthetic and natural) are usually the dominant fraction (Bagaev et al., 2017; Barrows et al., 2018; Gago et al., 2018; Woods et al., 2018).

Nowadays, fibers (macro and micro) are considered the most widespread anthropogenic particles in the world' oceans (Gago et al., 2018), but in spite of this, they have received little attention to date due to the difficulties to quantify and characterized them or avoid airborne contamination leading to an overestimation of the results (Rebelein et al., 2021).

On a global scale, the proportion of plastic fibers and microfibers in marine and surface water samples is about 70%; remaining 30% consists

\* Corresponding author.

E-mail address: [mdalessandro@ogs.it](mailto:mdalessandro@ogs.it) (M. D'Alessandro).

of natural fibers (Barrows et al., 2018). In the Mediterranean Sea, an important fraction of microfibers (35–72%) is represented by cotton, wool or cellulose (Musso et al., 2019; Pedrotti et al., 2020).

The spatial concentration of synthetic and natural fibers in the sea varies according to atmospheric and hydrodynamic conditions (Enders et al., 2015; Silvestrova & Stepanova, 2021; Wang et al., 2020). In general, higher concentrations are usually found in coastal areas than offshore (Desforges et al., 2014; Lusher et al., 2014; Manbohi, et al., 2021), probably influenced by the presence of urban settlements along the coast (Mathalon and Hill, 2014; Zhao et al., 2015). The concentration of synthetic fibers in the oceans also shows an increasing latitudinal gradient from north to south and the Mediterranean appears to be a large sink for this form of microplastic (Suaria et al., 2020). Furthermore, Silvestrova & Stepanova (2021) observed that fibers are usually concentrated in the subsurface layer, while microplastic particles remain concentrated in the surface layer. Conversely, Barrows et al. (2017) reported that most of the microplastics in superficial waters appear to be microfibers, with a length between 100  $\mu$ m and 5 mm and a width of about 1.5 orders of magnitude lower.

However, regardless of their spatial distribution, synthetic and natural fibers have been found in many marine species (vertebrates and invertebrates) (Avio et al., 2020; Bour et al., 2020; Collard et al., 2018; Compa et al., 2018; Halstead et al., 2018; Horton et al., 2017; Remy et al., 2015; Savoca et al., 2019; Silva-Cavalcanti et al., 2017) and this raise concern on the potential impacts they may have on animal communities and human health (Cox et al., 2019).

Thus, considering the ubiquitous presence and distribution and wide-ranging effects of microplastics and microfibers on marine organisms, there is an urgent need to develop long-term monitoring programs taking into account the different components (water, sediment and biota) of the marine ecosystem (Gallo et al., 2018) and, in particular, by selecting species with large distribution range and high abundances, sessile, easy to sample and tolerant to a wide range of environmental conditions (Beyer et al., 2017).

Barnacle crustaceans, due to their high tolerance to environmental stresses, are commonly used in marine pollution monitoring programs in coastal areas (Chen et al., 2015; Powell and White, 1990; Rainbow and Blackmore, 2001; Xu et al., 2020). In particular, barnacles of the genus *Lepas* (family Lepadidae) have a cosmopolitan distribution and are among the most abundant and widespread biofouling organisms globally (Thiel and Gutow, 2005), accounting for 90% of the biomass of the fouling community found on mooring systems (Martin et al., 2020). They are organisms that implant themselves on floating objects (Gil and Pfaller, 2016) and dominate assemblages over time (Astudillo et al., 2009; Goldstein et al., 2014), thanks to their greater resistance to predation due to the presence of hard plates (Iljin et al., 2013).

These barnacles are omnivorous (Barreiros and Teves, 2005; Mesaglio et al., 2021; Setsaas and Bester, 2006), feeding opportunistically on neustonic zooplankton, and any other organism in neuston (Bieri, 1966).

*L. (Lepas) anatifera* exhibits a coordinated behavior of cirri and mouthparts in the capture and ingestion of heterogeneous food of various sizes (Gravel, 1893). Food capture occurs through a synchronous movement between the cirri, which are very active in combing the surrounding water, and the peduncle that is very mobile and oscillates in all directions. The probability of obtaining food by casual contact with organisms is increased as the movement of the cirri results in foraging for food towards the mouth.

*L. (Lepas) anatifera* not only captures and feeds on large prey but it has a heterogeneous diet including crustaceans, polychaetes and molluscs. The feeding pattern can be traced to opportunistic rather than a selective behavior (Howard and Scott, 1958). Due to their feeding efficiency combined with their surface position in the water column and opportunistic behaviour they can be prone to the ingestion of microplastics and, therefore, they could be considered as suitable species to assess the presence and distribution of microplastics and fibers in the surrounding waters. Indeed, although the present study does not provide

data on the concentrations of microplastics in the environment, recent studies (Xu et al., 2020; Zhang et al., 2022) on the efficiency of some species of barnacles that have the same feeding habits of *Lepas anatifera* have highlighted the role of these species as bioindicators of surface and subsurface pollution by microplastics and microfibers. Indeed, thanks to their elongated shape, the fibers have the potential to be entangled in appendages, gill filaments and within the gastrointestinal system of numerous and heterogeneous organisms, and consequently they can harm them directly or cause negative physiological effects (Rebelein et al., 2021).

In this study, starting from the opportunistic feeding strategy and the subsurface position of *L. (Lepas) anatifera*, adhered to the several substrates, including buoys, we aim validating the hypothesis that this species is able to ingest floating microplastics and microfibers and to suggest the use of *L. (Lepas) anatifera* in assessing the presence of microplastics and micro fibers in coastal marine waters.

## 2. Materials and methods

### 2.1. Sampling area

The study area is the Capo Milazzo MPA located in the north-eastern sector of Sicily (Fig. 1). Having been established in 2019, it is the youngest MPA established in Italy.

The promontory of Capo Milazzo extends approximately 6 km offshore and, due to its conformation, represents a separating element between the Gulf of Patti and the Gulf of Milazzo. Both are highly urbanized areas and receive significant contributions from urban and industrial wastewater due to the presence of an important industrial pole in Milazzo, which in 2005 has been declared as a Contaminated Site of National Interest (SIN) (D'Agostino et al., 2020). The pressure of plastic pollution along the beaches is strong and is often associated with the transport of large quantities of waste from rivers during rainfall events.

### 2.2. *Lepas (Lepas) anatifera* sampling

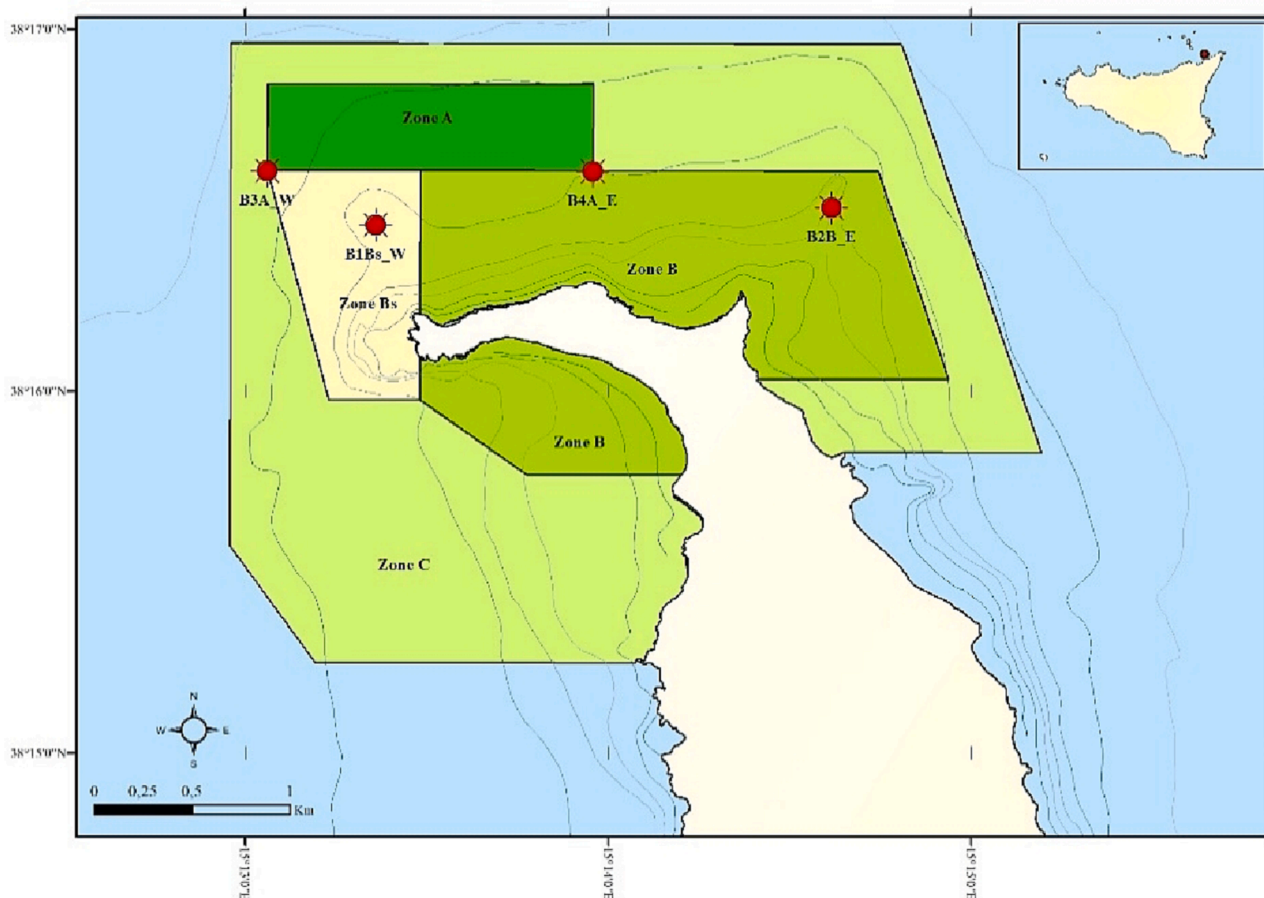
Between April and May 2021, individuals of *L. (Lepas) anatifera* were sampled at 4 sites within the Capo Milazzo MPA. Sampling was carried out on 2 buoys installed in July 2020 for the mooring of diving boats on the Secca di Ponente (B1Bs\_W) (BS zone - special general reserve) and Secca di Levante (B2B\_E) (Zone B - general reserve). The two buoys are respectively about 0.25 and 0.30 nm from the coast and 0.97 nm between them. The other 2 sampling buoys, also positioned in July 2020, delimit zone A (B3A\_W and B4A\_E) (integral reserve) and are 0.7 nm and 0.37 nm respectively from the coast and 0.60 nm between them (Table 1).

Under each buoy, 30 individuals of *L. (Lepas) anatifera* were collected and immediately frozen at  $-20^{\circ}$  C, for a total of 120 individuals (Fig. 2).

### 2.3. Microplastics analysis

In the laboratory, the length (mm) of capitulum and peduncle of each individual were measured, and the stomach and intestinal tracts were taken. Then, the Gastro-intestinal tract (GIT) of each individual was weighed, opened and placed in glass beakers at 1:3 (w/v) with 10% KOH. The solutions were incubated in an oven at  $60 \pm 5^{\circ}$  C for 6 h and subsequently left at room temperature overnight. Samples were filtered through glass fiber filters (1.6  $\mu$ m pore size, Whatman GF / A, GE Healthcare, UK) using a vacuum pump (Schirinzi et al., 2020).

The filters obtained from the digestion of the GIT samples were examined under a Zeiss Discovery V.8 stereo microscope coupled with AxioVision digital image processing software. All particles recovered on the filter were photographed, counted and measured (length and width). Subsequently, they were classified based on their shape (fibers, fragments, tangled fibers and sheets) and color (Galgani et al., 2013). Fibers



**Fig. 1.** Study area. The red dots represent the sampling sites and green polygons the different protection zones (A, B, C) of the Capo Milazzo MPA. (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article.)

**Table 1**  
Sampling sites of *L. (Lepas) anatifera* and level of protection of the area.

Protection level	Area A	Latitude	Longitude
Integral reserve (Zone A)	B3A_W	38,2768583333	15,2176666667
		N	E
Integral reserve (Zone A)	B4A_E	38,2768277778	15,2326305556
		N	E
Special general reserve (Zone Bs)	B1Bs_W	38,2743610422	15,2226331199
		N	E
General reserve (Zone B)	B2B_E	38,2751525914	15,2435842183
		N	E

with length between 100  $\mu\text{m}$  and 5 mm were classified as “microfibers”, all fibers produced by chemical synthesis of petroleum products were classified as “synthetic microfibers”, while fibers from natural materials not chemically synthesized (cotton, wool, cellulose) were classified as “natural microfibers”. A taxonomic identification of organisms with shells and valves that did not undergo the digestion process with KOH was also performed.

All procedures were carried out following strict measures to avoid airborne contamination of the samples and using filter papers in Petri dishes exposed to the laboratory air, as control blanks during analysis (Giani et al., 2019; Lusher et al., 2017; Manbohi et al., 2021; Xu et al., 2020).

Fourier transform infrared (FTIR) spectroscopy technique was used to identify polymers nature in *L. anatifera* samples using a Cary 630 FTIR Spectrometer (Agilent) and analyzed using the software Micro Lab FTIR (Agilent). Only spectra with a percentage of matching with the spectra in the library > 80% were considered in the analysis,

Pearson’s test was applied to evaluate possible correlations between the morphological data of the crustacean’s and the fiber length. In addition, the recorded fragments ingestion density was calculated for each sampling area (R version 4.0.4 packages PerformanceAnalytics, ggplot2; hrbrthemes; dplyr; tidy; viridis).

### 3. Results

Thirty-six out of a total of 120 individuals analysed (30%) have fibers and fragments within the digestive tract (Table 2).

In only one individual, fibers were attached to the cirri (Fig. 3).

The highest recovery percentages were found in *L. (Lepas) anatifera* collected in stations B4A\_E and B1Bs\_W (60%), followed by that in B3A\_W (46%) and B2B\_E (43.3%). 7.4% (8 ind.) had more than one fiber type and 18.5% (20 ind.) ingested more than one fiber. The highest density of organisms that ingested multiple fibers was recorded at station B1Bs\_W, while the highest density of organisms that did not ingest fibers was found in station B2B\_E (Fig. 4).

Overall, plastic particles were found within the GIT of *L. anatifera* in two different shapes: fibers (85.9%) and fragments (14.1%). In four specimens, the fibers were twisted to form a tangle (3.3% of specimens) (Fig. 5).

Due to their small size, 57.7% of the extracted fibers were indeterminate, as they did not show an identifiable spectrum. Regarding the other microfibers, Fig. 6 shows their qualitative composition obtained by FTIR analysis: 30.6% synthetic fibers (Polyamide, Nylon, Polyvinyl chloride, Polyethylene) and 11.7% natural fibers (cellulose and cotton) (Fig. 6). The predominant colour is transparent (47.2%) followed by blue (25%) and white (9.26%) (Fig. 6). The fibers appear isolated and





Fig. 2. *L. (Lepas) anatifera* attached to one of the sampling buoy.

Table 2

Percentage of fibers/fragments found in *L. (Lepas) anatifera*.

	number	%
Tot individuals	120	
Ind. with micro fibers/fragment	36	30.0
Ind without fibers	84	70.0
Synthetic fibers	34	30.6
Tot fibers	111	
Indeterminate fibers	64	57.7
Number of natural fibers	13	11.7
Individuals who have ingested multiple types of fiber	8	6.7
Individuals who have ingested more than 1 fiber	20	16.7

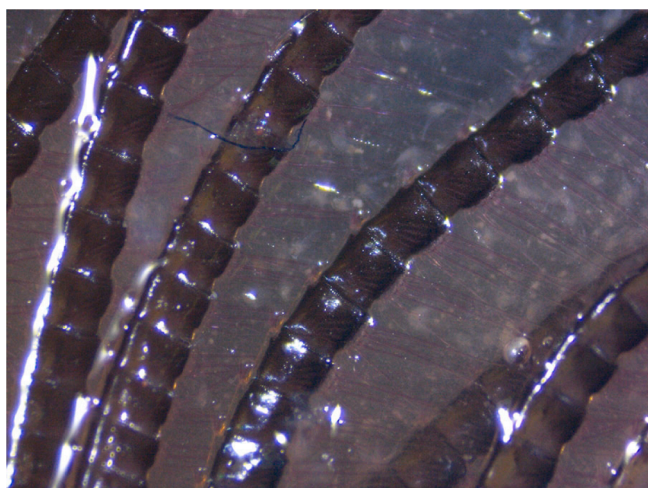


Fig. 3. Fibers attached to the cirri of an individual of *L. (Lepas) anatifera*.

only in one case intertwined.

Regarding the size of the particles, overall, the main percent abundance of fibers ranged between  $\geq 1$  and  $< 2$  mm (33.3%), followed by fibers comprised between 0 and 1 mm (29.6%) and  $\geq 2$  and  $< 3$  mm. Only 1.9% of the identified fibers were bigger than 5 mm. No significant correlations were found between the length of the fragments in the GIT of *L. (Lepas) anatifera* and the morphological parameters of the examined organisms (Fig. 7). Average number of polymers and natural fibers was  $1.74 \pm 0.80$  items/individual.

Other than the plastic particles, the qualitative analysis on the

undigested material has shown that *Lepas anatifera* can access heterogeneous food resources. In the study area, indeed, the availability of food is high and the diet of the individuals examined consists mainly of Ostracoda, and larval shell of pelagic mollusc *Oxygyrus* sp. (fam. Atlantidae) and *Creseis* spp. (fam. Creseidae) (Fig. 8).

#### 4. Discussion

Sessile organisms such as *L. (Lepas) anatifera* are the most common component of marine fouling. Their life as “suspension feeders” leads them to consume a wide variety of suspended food particles, including small zooplankton (Ye & Andrady 1991, Tsikhon-Lukanina et al., 2001), and their survival is linked to physical (light, temperature) and biological (food availability) processes (Inatsuchi et al., 2010). The present study, carried out on *L. (Lepas) anatifera* sampled on a fixed mooring system, provides a precise estimate of the efficiency of these biofoulers in capturing microplastic particles and microfibers floating in the water column. Considering that the buoys were positioned in July 2020, it is possible to state that the individuals sampled come from larvae that settled immediately after the buoys were installed and therefore the maximum age for larger individuals is  $< 1$  year since the sampling was performed in April and May 2021.

Within the Capo Milazzo MPA, *L. (Lepas) anatifera* is only observed under the buoys in areas exposed to strong currents in the northern sector of the promontory bordering the Gulf of Patti. The presence of microfibers in the 30% of the total 120 individuals examined coincides with other surveys of *L. (Lepas) anatifera* and *L. pacifica* in the north Pacific area, where 33.5% of individuals ingested microplastic (Goldstein and Goodwin, 2013).

Our results revealed that in the four sampling sites the greater concentration of fibres in *L. (Lepas) anatifera* is recorded in the western sector of the MPA (B1Bs, B3A\_W), while *L. (Lepas) anatifera* in the eastern sector seems to be less impacted by the presence of micro fibers and / or fragments. The western sector of the area, due to the hydrological conditions, is more exposed to the contribution of organic matter coming from the several urban wastewaters of the adjacent Gulf of Patti, which could be among the most responsible for the presence of fibers, microfibrils and microplastics in the water column.

However, there is little information on the presence of microplastics in the waters of the study area. A study, conducted by Savoca et al (2020) in the nearby Gulf of Patti, reported the presence of microplastics and artificial microfibers in the gastrointestinal tract of late larval and juvenile stages of clupeid fishes *Sardina pilchardus* (0.53 items/sample) and *Engraulis encrasicolus* (0.26 items/sample). According to our results,



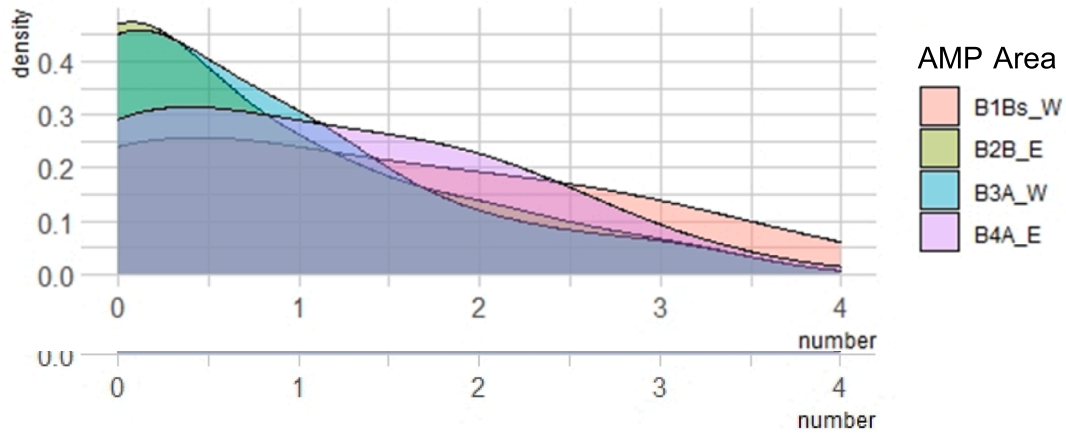


Fig. 4. Comparison of anthropogenic particles densities found in each sampling area.

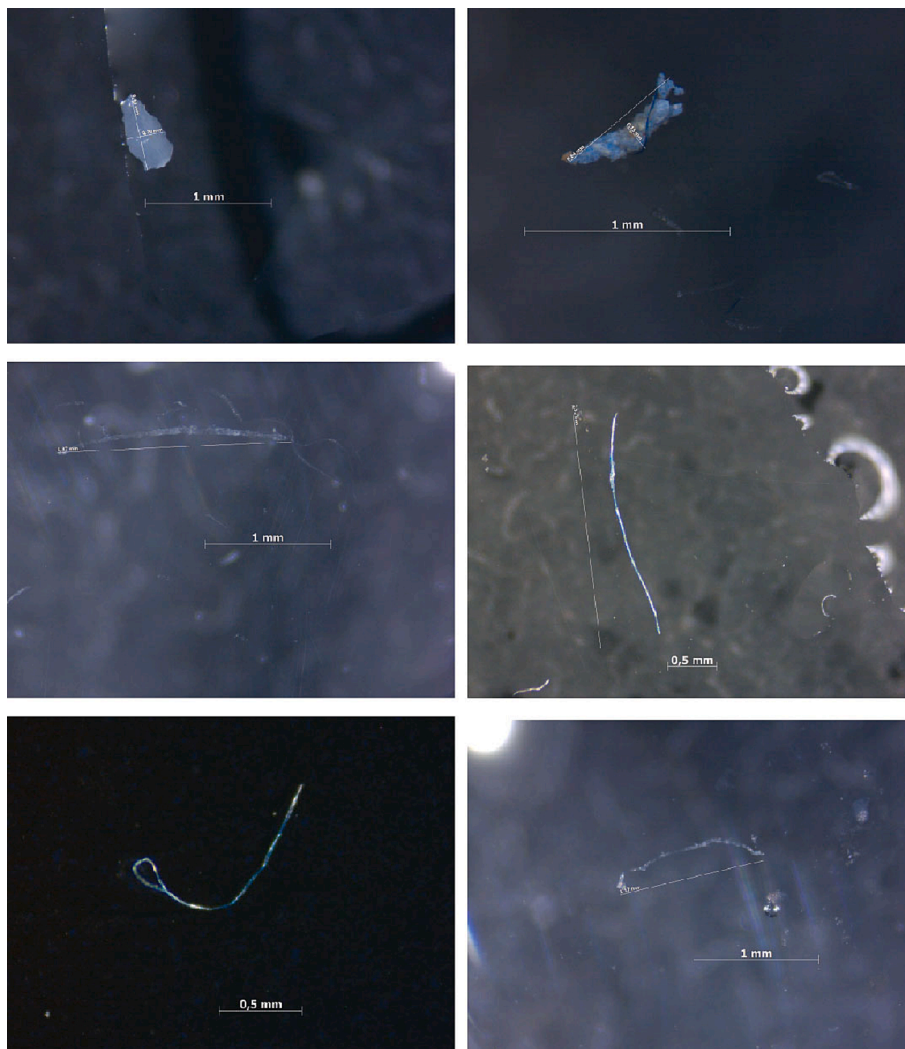


Fig. 5. Examples of extracted MPs (fragments and fibers),

these concentrations are lower than that observed in the GIT of *Lepas anatifera*. However, the typology of polymers found in our study seems to match with those recorded by Savoca et al. (2020): polyester, polypropylene, polyacrylonitrile, polyethylene, polyamide, nylon, rayon, and polyurethane. Moreover, in the same area Savoca et al. (2019) reported the presence of artificial cellulose microfibrils also in 19

individuals out of a sample of 30 individuals (63.3%) of another fish species, *Boops boops*, and for a total of 80 fibres. These differences could be due to both the high volume of tourists and the presence of numerous streams that exist on the Gulf of Patti and are capable of discharging large amounts of debris into the sea, including possibly plastics and land waste, especially during flash floods (Savoca et al., 2019; Pierdomenico

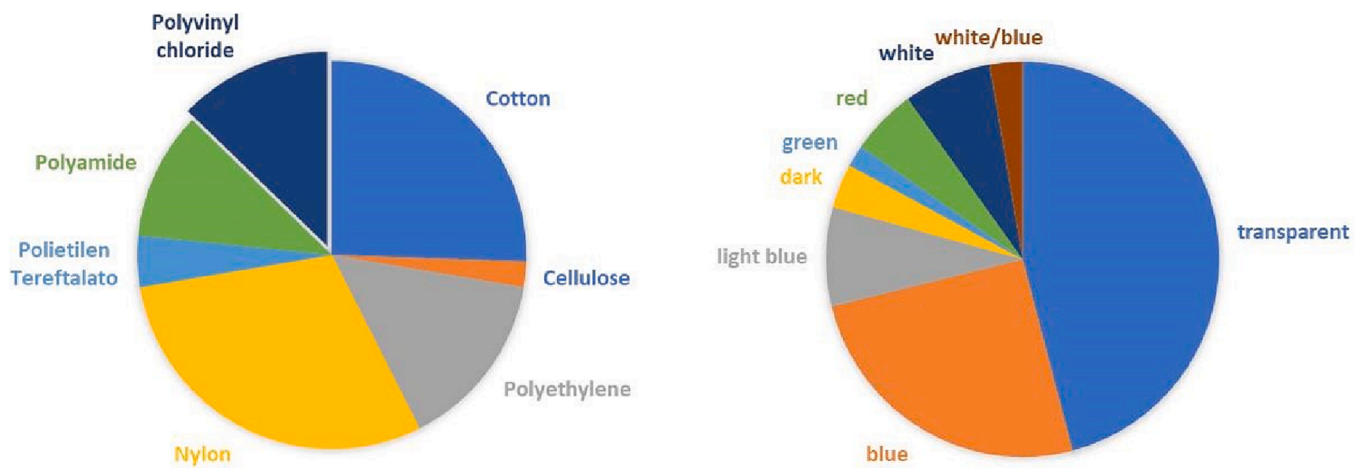


Fig. 6. Composition of polymers and natural fibers (right) and percentage of colour typology of fragments (left) found in *L. (Lepas) anatifera* GIT.

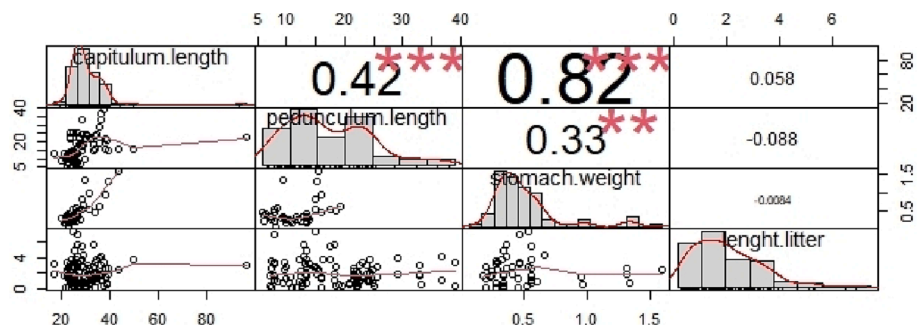


Fig. 7. Pearson correlation matrix plot of morphological parameters (stomach. weight, capitulum and pedunculum length) and length of fragments. Bold data represent  $P < 0.05$ ; \*,  $P < 0.01$ ; \*\*,  $P < 0.001$ ; \*\*\*.

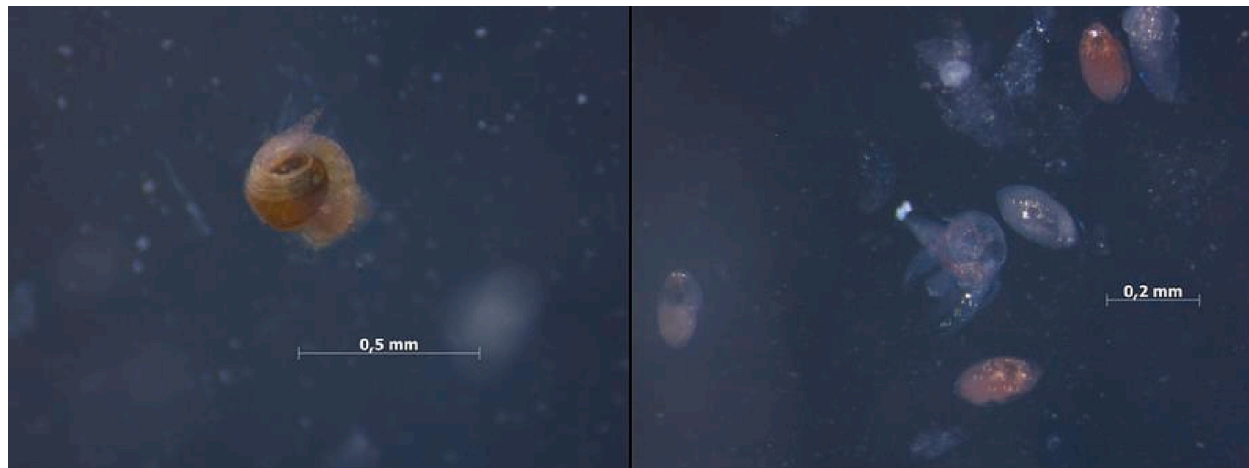


Fig. 8. Prey found in GITs of *L. (Lepas) anatifera*: a) Larval shell of pelagic molluscs *Creseis* sp., b) Ostracoda.

et al., 2022).

In addition, due to their position along the air-sea and subsurface interface, barnacles are likely to encounter floating microplastics more frequently than pelagic and mesopelagic fish which usually perform vertical migrations along the water column. In addition, the mean numbers of plastics recorded in our study is greater than those observed in *L. anatifera* in Galapagos Island ( $0.71 \pm 0.29$ ) (Jones et al., 2021) and barnacles *Balanus glandula* ( $1.2 \pm 1.9$ ) in British Columbia, Canada (Davies et al., 2021).

Most of the microfibrils identified during the present study, due to

their size, were found to be below the detection limit of the FT-IR Spectrometer used during the analyses.

However, the wide variety, by type, shape and color, of fibers and plastic fragments that we found in individuals of *L. (Lepas) anatifera* together with the high abundances of mollusc and ostracod shells, confirm the non-food selectivity of these organisms. *Lepas anatifera* can indifferently capture any type of prey (Goldstein and Goodwin, 2013) and therefore are unable to differentiate between microplastics and prey. For this reason, it is necessary to encourage studies on the habits of many invertebrates in capturing microplastics and fibers in natural

conditions and not only under controlled conditions and to better understand the role of biological interactions within the community in sequestering the fraction of microplastics present in the oceans and to determine the levels of contamination to which they are subject.

However, as highlighted by several authors, the toxicological effects and the implications on fitness of exposure to microplastics are more consistent in all those animals that occupy the lowest trophic levels. Filter-feeding crustaceans, such as copepoda, amphipoda and cirripeda are more exposed to the consumption of microplastics and micro-fibers (Foley et al., 2018; Setälä et al., 2016; Walkinshaw et al., 2020).

## 5. Conclusion

Few studies have been conducted on microplastics in barnacles, but the results are suggesting the ability of these organisms to ingest microplastics (Zhang et al., 2021). According to current research results, uptake of microplastics by barnacles is ubiquitous, and the degree of pollution of the habitat is the main influencing factors (Gajahin et al., 2017; Xu et al., 2020). This relationship has already been highlighted in other studies, in Mediterranean and extra-Mediterranean contexts who results highlighter as synthetic fibers are more present along the coasts, on a global scale, with higher concentrations in densely populated areas and where there are urban treated wastewaters (Browne et al., 2011; Henry et al., 2019; Pedrotti et al., 2020).

Furthermore, belonging to different ecotypes seems to determine the mechanism of microplastics accumulation in barnacles (Xu et al., 2020). Among barnacles, *L. anatifera* is the one that ingests the largest dimensions of microplastics and microfibrils (Goldstein, 2013). Probably due to its larger body size or feeding strategy, but this aspect needs further confirmation.

However, the retention and egestion times of ingested microplastics and microfibrils in barnacles are not known and we are not able to directly compare the concentrations of microplastics in the surface waters in which the *L. anatifera* samples were collected. Our finding evidenced that *L. (Lepas) anatifera*, and barnacles in general, thanks to its opportunist feeding behaviour and the consequent ingestion of microplastic particles, to its abundance in macrofouling assemblages, wide distribution and to the ease of sampling, should be used in the assessment of the ecological status of pelagic and coastal ecosystems as already pointed out also in other areas of the world (Xu et al., 2020). It is therefore necessary to work to provide data to select the most suitable organisms for monitoring the presence of microplastics, emphasizing that most living organisms, both vertebrates and invertebrates, come into contact, directly or indirectly, with these particles which can cause harmful effects throughout the food chain.

## CRedit authorship contribution statement

**Gianfranco Scotti:** Conceptualization, Methodology, Investigation, Resources, Writing – original draft. **Michela D'Alessandro:** Formal analysis, Writing – review & editing, Supervision. **Valentina Esposito:** Formal analysis, Writing – review & editing, Supervision. **Pietro Vivona:** Data curation, Writing – review & editing. **Cristina Panti:** Conceptualization, Methodology, Supervision, Investigation, Writing – review & editing.

## Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

## Data availability

Data will be made available on request.

## Acknowledgements

We would like to thank the staff, coordinators and volunteers of Capo Milazzo Marine Protected Area and in particular to Peppe Maimone and Giovanni Mangano for their contribution during the field surveys.

## References

- Astudillo, J.C., Bravo, M., Dumont, C.P., Thiel, M., 2009. Detached aquaculture buoys in the SE Pacific: potential dispersal vehicles for associated organisms. *Aquat. Biol.* 5, 219–231. <https://doi.org/10.3354/ab00151>.
- Avio, C.G., Pittura, L., d'Errico, G., Abel, S., Amorello, S., Marino, G., Gorbi, S., Regoli, F., 2020. Distribution and characterization of microplastic particles and textile microfibrils in Adriatic food webs: general insights for biomonitoring strategies. *Environ. Pollut.* 258, 113766 <https://doi.org/10.1016/j.envpol.2019.113766>.
- Bagaev, A., Mizyuk, A., Khatmullina, L., Isachenko, I., Chubarenko, I., 2017. Anthropogenic fibres in the Baltic Sea water column: Field data, laboratory and numerical testing of their motion. *Sci. Total Environ.* 599, 560–571. <https://doi.org/10.1016/j.scitotenv.2017.04.185>.
- Barreiros, J.P., Teves, M., 2005. The sunfish *Mola mola* as an attachment surface for the lepadid cirriped *Lepas anatifera*: a previously unreported association. *qua-Internacional. J. Ichthyol.* 10 (1), 1–4.
- Barrows, A.P.W., Cathey, S.E., Petersen, C.W., 2018. Marine environment microfiber contamination: global patterns and the diversity of microparticle origins. *Environ. Pollut.* 237, 275–284.
- Barrows, A.P., Neumann, C.A., Berger, M.L., Shaw, S.D., 2017. Grab vs. neuston tow net: a microplastic sampling performance comparison and possible advances in the field. *Anal. Methods* 9 (9), 1446–1453.
- Beyer, J., Green, N.W., Brooks, S., Allan, I.J., Ruus, A., Gomes, T., Brate, I.L.N., Schøyen, M., 2017. Blue mussels (*Mytilus edulis* spp.) as sentinel organisms in coastal pollution monitoring: a review. *Mar. Environ. Res.* 130, 338–365. <https://doi.org/10.1016/j.marenvres.2017.07.024>.
- Bieri, R., 1966. Feeding preferences and rates of the snail, *Ianthina prolongata*, the barnacle, *Lepas anatifera*, the nudibranchs, *Glaucus atlanticus* and *Fiona pinnata*, and the food web in the marine neuston. *Publ. Seto Marine Biol. Labor.* 14 (2), 161–170.
- Bour, A., Hossain, S., Taylor, M., Sumner, M., Carney Almroth, B., 2020. Synthetic microfiber and microbead exposure and retention time in model aquatic species under different exposure scenarios. *Front. Environ. Sci.* 8, 83. <https://doi.org/10.3389/fenvs.2020.00083>.
- Browne, M.A., Crump, P., Niven, S.J., Teuten, E., Tonkin, A., Galloway, T., Thompson, R., 2011. Accumulation of microplastic on shorelines worldwide: sources and sinks. *Environ. Sci. Tech.* 45 (21), 9175–9179. <https://doi.org/10.1021/es201811n>.
- Burns, E.E., Boxall, A.B.A., 2018. Microplastics in the aquatic environment: evidence for or against adverse impacts and major knowledge gaps. *Environ. Toxicol. Chem.* 37 (11), 2776–2796. <https://doi.org/10.1002/etc.4268>.
- Chen, L., Lam, J.C., Zhang, X., Pan, K., Guo, C., Lam, P.K., Wang, W., Liu, H., Qian, P.Y., 2015. Relationship between metal and polybrominated diphenyl ether (PBDE) body burden and health risks in the barnacle *Balanus amphitrite*. *Mar. Pollut. Bull.* 100 (1), 383–392. <https://doi.org/10.1016/j.marpolbul.2015.08.020>.
- Collard, F., Gasperi, J., Gilbert, B., Eppe, G., Azimi, S., Rocher, V., Tassin, B., 2018. Anthropogenic particles in the stomach contents and liver of the freshwater fish *Squalius cephalus*. *Sci. Total Environ.* 643, 1257–1264. <https://doi.org/10.1016/j.scitotenv.2018.06.313>.
- Compa, M., Ventero, A., Iglesias, M., Deudero, S., 2018. Ingestion of microplastics and natural fibres in *Sardina pilchardus* (Walbaum, 1792) and *Engraulis encrasicolus* (Linnaeus, 1758) along the Spanish Mediterranean coast. *Mar. Pollut. Bull.* 128, 89–96. <https://doi.org/10.1016/j.marpolbul.2018.01.009>.
- Cox, K.D., Covernton, G.A., Davies, H.L., Dower, J.F., Juanes, F., Dudas, S.E., 2019. Human consumption of microplastics. *Environ. Sci. Tech.* 53 (12), 7068–7074. <https://doi.org/10.1021/acs.est.9b01517>.
- D'Agostino, F., Bellante, A., Quinci, E., Gherardi, S., Placenti, F., Sabatino, N., Buffa, G., Avellone, G., Di Stefano, V., Del Core, M., 2020. Persistent and Emerging Organic Pollutants in the Marine Coastal Environment of the Gulf of Milazzo (Southern Italy): Human Health Risk Assessment. *Front. Environ. Sci.* 8, 117. <https://doi.org/10.3389/fenvs.2020.00117>.
- Desforges, J.P.W., Galbraith, M., Dangerfield, N., Ross, P.S., 2014. Widespread distribution of microplastics in subsurface seawater in the NE Pacific Ocean. *Mar. Pollut. Bull.* 79 (1–2), 94–99. <https://doi.org/10.1016/j.marpolbul.2013.12.035>.
- Enders, K., Lenz, R., Stedmon, C.A., Nielsen, T.G., 2015. Abundance, size and polymer composition of marine microplastics >=10µm in the Atlantic Ocean and their modelled vertical distribution. *Mar. Pollut. Bull.* 100 (1), 70–81. <https://doi.org/10.1016/j.marpolbul.2015.09.027>.
- Foley, C.J., Feiner, Z.S., Malinich, T.D., Höök, T.O., 2018. A meta-analysis of the effects of exposure to microplastics on fish and aquatic invertebrates. *Sci. Total Environ.* 631, 550–559. <https://doi.org/10.1016/j.scitotenv.2018.03.046>.
- Gago, J., Carrtero, O., Filgueiras, A.V., Viñas, L., 2018. Synthetic microfibrils in the marine environment: a review on their occurrence in seawater and sediments. *Mar. Pollut. Bull.* 127, 365–376.
- Gajahin, G.N.T., Jayan, D.M.S., Amaratne, Y., Suchana, C., 2017. Effects of Microplastics on barnacles and wild bivalves in the eastern coast of Thailand: an approach to coastal zone conservation. *Mar. Pollut. Bull.* 124, 349–355. <https://doi.org/10.1016/j.marpolbul.2017.06.010>.



- Galgani, F., Hanke, G., Werner, S.D.V.L., De Vrees, L., 2013. Marine litter within the European marine strategy framework directive. *ICES J. Mar. Sci.* 70 (6), 1055–1064.
- Gallo, F., Fossi, C., Weber, R., Santillo, D., Sousa, J., Ingram, I., Nadal, A., Romano, D., 2018. Marine litter plastics and microplastics and their toxic chemicals components: the need for urgent preventive measures. *Environ. Sci. Eur.* 30 (1), 1–14.
- Giani, D., Baimi, M., Galli, M., Casini, S., Fossi, M.C., 2019. Microplastics occurrence in edible fish species (*Mullus barbatus* and *Merluccius merluccius*) collected in three different geographical sub-areas of the Mediterranean Sea. *Mar. Pollut. Bull.* 140, 129–137. <https://doi.org/10.1016/j.marpolbul.2019.01.005>.
- Gil, M.A., Pfaller, J.B., 2016. Oceanic barnacles act as foundation species on plastic debris: implications for marine dispersal. *Sci. Rep.* 6, 1–7.
- Goldstein, M.C., Goodwin, D.S., 2013. Gooseneck barnacles (*Lepas* spp.) ingest microplastic debris in the North Pacific Subtropical Gyre. *PeerJ* 1, e184.
- Goldstein, M.C., Carson, H.S., Eriksen, M., 2014. Relationship of diversity and habitat area in North Pacific plastic-associated rafting communities. *Mar. Biol.* 161 (6), 1441–1453.
- Gruvel, A., 1893. Contributions a l'étude des cirrhipèdes. *Arch. Zool. exp. gen., Sér. 3. T. 1*, 401–610.
- Gunaalan, K., Fabbri, E., Capolupo, M., 2020. The hidden threat of plastic leachates: a critical review on their impacts on aquatic organisms. *Water Res.* 184, 116170.
- Hahladakis, J.N., Velis, C.A., Weber, R., Iacovidou, E., Purnell, P., 2018. An overview of chemical additives present in plastics: migration, release, fate and environmental impact during their use, disposal and recycling. *J. Hazard. Mater.*, 344:179–199. <https://doi.org/10.1016/j.jhazmat.2017.10.014>.
- Halstead, J.E., Smith, J.A., Carter, E.A., Lay, P.A., Johnston, E.L., 2018. Assessment tools for microplastics and natural fibres ingested by fish in an urbanised estuary. *Environ. Pollut.* 234, 552–561. <https://doi.org/10.1016/j.envpol.2017.11.085>.
- Henry, B., Laitala, K., Klepp, I.G., 2019. Microfibres from apparel and home textiles: prospects for including microplastics in environmental sustainability assessment. *Sci. Total Environ.* 652, 483–494. <https://doi.org/10.1016/j.scitotenv.2018.10.166>.
- Horton, A.A., Walton, A., Spurgeon, D.J., Lahive, E., Svendsen, C., 2017. Microplastics in freshwater and terrestrial environments: evaluating the current understanding to identify the knowledge gaps and future research priorities. *Sci. Total Environ.* 586, 127–141. <https://doi.org/10.1016/j.scitotenv.2017.01.190>.
- Howard, G.K., Scott, H.C., 1959. Predaceous feeding in two common gooseneck barnacles. *Science* 129 (3350), 717–718. <https://doi.org/10.1126/science.129.3350.717>.
- Ilijin, I.N., Petrosyan, V.G., Bessonov, S.A., Dergunova, N.N., 2013. Modeling of the invasion and development of the pelagic communities of fouling organisms in the ocean. *Russ. J. Biol. Invasions* 4 (4), 225–233. <https://doi.org/10.1134/S20751171304005X>.
- Inatsuchi, A., Yamato, S., Yusa, Y., 2010. Effects of temperature and food availability on growth and reproduction in the neustonic pedunculate barnacle *Lepas anserifera*. *Mar. Biol.* 157 (4), 899–905. <https://doi.org/10.1007/s00227-009-1373-0>.
- Jones, J.S., Guézou, A., Medor, S., Nickson, C., Savage, G., Alarcón-Ruales, D., Lewis, C., 2022. Microplastic distribution and composition on two Galápagos island beaches, Ecuador: Verifying the use of citizen science derived data in long-term monitoring. *Environ. Pollut.* 311.
- Lima, A.R.A., Ivar do Sul, J.A., Frias, J.P.G.L., Panti, C. (Eds.), 2021. Microplastics in the Marine Environment: Sources, Distribution, Biological Effects and Socio-Economic Impacts. *Lausanne: Frontiers Media SA*. doi: 10.3389/978-2-88966-871-7.
- Lusher, A.L., Burke, A., O'Connor, I., Officer, R., 2014. Microplastic pollution in the Northeast Atlantic Ocean: validated and opportunistic sampling. *Mar. Pollut. Bull.* 88 (1–2), 325–333. <https://doi.org/10.1016/j.marpolbul.2014.08.023>.
- Lusher, A.L., Welden, N.A., Sobral, P., Cole, M., 2017. Sampling, isolating and identifying microplastics ingested by fish and invertebrates. *Anal. Methods* 9 (9), 1346–1360.
- Manbohi, A., Mehdiinia, A., Rahnama, R., Dehbandi, R., 2021. Microplastic pollution in inshore and offshore surface waters of the southern Caspian Sea. *Chemosphere* 281, 130896. <https://doi.org/10.1016/j.chemosphere.2021.130896>.
- Martin, M.V., Venkatesan, R., Beyline, M., Limna, M.V.P., Divya, L., 2020. Influence of environmental factors on macrofouling assemblages on moored buoys in the eastern Arabian Sea. *PLoS One* 15 (1). <https://doi.org/10.1371/journal.pone.0223560>.
- Mathalon, A., Hill, P., 2014. Microplastic fibers in the intertidal ecosystem surrounding Halifax Harbor, Nova Scotia. *Mar. Pollut. Bull.* 81 (1), 69–79. <https://doi.org/10.1016/j.marpolbul.2014.02.018>.
- Mesaglio, T.P., Schilling, H.T., Adler, L., Ahyong, S.T., Maslen, B., Suthers, I.M., 2021. The ecology of *Lepas*-based biofouling communities on moored and drifting objects, with applications for marine forensic science. *Mar. Biol.* 168 (2), 1–16. <https://doi.org/10.1007/s00227-021-03822-1>.
- Musso, M., Achtypi, A., Bassotto, D., Suaria, G., Aliani, S., 2019. Textile microfibers in Mediterranean surface waters. *CIESM-Congress*.
- Pedrotti, M.L., Petit, S., Eyheraguibel, B., Kerros, M.E., Elineau, A., Ghiglione, J.F., Loret, J.F., Rostan, A., Gorsky, G., 2020. Pollution by anthropogenic microfibers in North-West Mediterranean Sea and efficiency of microfiber removal by a wastewater treatment plant. *Sci. Total Environ.* 758, 144195. <https://doi.org/10.1016/j.scitotenv.2020.144195>.
- Pierdomenico, M., Ridente, D., Casalbone, D., Di Bella, L., Milli, S., Chiocci, F.L., 2022. Plastic burial by flash-flood deposits in a prodelta environment (Gulf of Patti, Southern Tyrrhenian Sea). *Mar. Pollut. Bull.* 181, 113819. <https://doi.org/10.1016/j.marpolbul.2022.113819>.
- Powell, M.I., White, K.N., 1990. Heavy metal accumulation by barnacles and its implications for their use as biological monitors. *Mar. Environ. Res.* 30 (2), 91–118. [https://doi.org/10.1016/0141-1136\(90\)90013-E](https://doi.org/10.1016/0141-1136(90)90013-E).
- Rainbow, P.S., Blackmore, G., 2001. Barnacles as biomonitors of trace metal availabilities in Hong Kong coastal waters: changes in space and time. *Mar. Environ. Res.* 51 (5), 441–463. [https://doi.org/10.1016/S0141-1136\(00\)00254-3](https://doi.org/10.1016/S0141-1136(00)00254-3).
- Rebelein, A., Int-Veen, I., Kammann, U., Scharssack, J.P., 2021. Microplastic fibers—underestimated threat to aquatic organisms? *Sci. Total Environ.* 777, 146045.
- Remy, F., Collard, B., Gilbert, P., Compoère, G., Eppe, G., Lepoint, G., 2015. When microplastic is not plastic: The ingestion of artificial cellulose fibers by macrofauna living in seagrass macrophytoderitrus. *Environ. Sci. Tech.* 49, 11158–11166. <https://doi.org/10.1021/acs.est.5b02005>.
- Savoca, S., Capillo, G., Mancuso, M., Faggio, C., Panarello, G., Crupi, R., Bonsignore, M., D'Urso, L., Compagnini, G., Neri, F., Fazio, E., Romeo, T., Bottari, T., Spanò, N., 2019. Detection of artificial cellulose microfibrils in *Boops boops* from the northern coasts of Sicily (Central Mediterranean). *Sci. Total Environ.* 691, 455–465. <https://doi.org/10.1016/j.scitotenv.2019.07.148>.
- Savoca, S., Bottari, T., Fazio, E., Bonsignore, M., Mancuso, M., Luna, G.M., Spanò, N., 2020. Plastics occurrence in juveniles of *Engraulis encrasicolus* and *Sardina pilchardus* in the Southern Tyrrhenian Sea. *Sci. Total Environ.* 718, 137457.
- Schirinzi, G.F., Pedà, C., Battaglia, P., Laface, F., Galli, M., Baimi, M., Consoli, P., Scotti, G., Esposito, V., Faggio, C., Farrè, M., Barcelo', D., Fossi, M.C., Andaloro, F., Romeo, T., 2020. A new digestion approach for the extraction of microplastics from gastrointestinal tracts (GITs) of the common dolphinfish (*Coryphaena hippurus*) from the western Mediterranean Sea. *J. Hazard. Mater.* 397, 122794. <https://doi.org/10.1016/j.jhazmat.2020.122794>.
- Setälä, O., Norkko, J., Lehtiniemi, M., 2016. Feeding type affects microplastic ingestion in a coastal invertebrate community. *Mar. Pollut. Bull.* 102 (1), 95–101. <https://doi.org/10.1016/j.marpolbul.2015.11.053>.
- Setsaas, T.H., Bester, M.N., 2006. Goose barnacle (*Lepas australis*) infestation of the Subantarctic fur seal (*Arctocephalus tropicalis*). *Afr. Zool.* 41 (2), 305–307. <https://doi.org/10.1080/15627020.2006.11407368>.
- Sharma, S., Sharma, V., Chatterjee, S., 2021. Microplastics in the Mediterranean Sea: sources, pollution intensity, sea health, and regulatory policies. *Front. Mar. Sci.* 8, 634934. <https://doi.org/10.3389/fmars.2021.634934>.
- Silva-Cavalcanti, J.S., Silva, J.D.B., de França, E.J., de Araújo, M.C.B., Gusmao, F., 2017. Microplastics ingestion by a common tropical freshwater fishing resource. *Environ. Pollut.* 221, 218–226. <https://doi.org/10.1016/j.envpol.2016.11.068>.
- Silvestrova, K., Stepanova, N., 2021. The distribution of microplastics in the surface layer of the Atlantic Ocean from the subtropics to the equator according to visual analysis. *Mar. Pollut. Bull.* 162, 111836. <https://doi.org/10.1016/j.marpolbul.2020.111836>.
- Suaria, G., Achtypi, A., Perold, V., Lee, J.R., Pierucci, A., Borrmann, T.G., Aliani, S., Ryan, P.G., 2020. Microfibers in oceanic surface waters: A global characterization. *Sci. Adv.* 6 (23), eaay8493. <https://doi.org/10.1126/sciadv.aay8493>.
- Teuten, E.L., Saquing, J.M., Knappe, D.R.U., Barlaz, M.A., Jonsson, S., Björn, A., Rowland, S.J., Thompson, R.C., Galloway, T.S., Yamashita, R., Ochi, D., Watanuki, Y., Moore, C., Viet, P.H., Tana, T.S., Prudente, M., Boonyatumanond, R., Zakaria, M.P., Akhavan, K., Ogata, Y., Hirai, H., Iwasa, S., Mizukawa, K., Hagino, Y., Imamura, A., Saha, M., Takada, H., 2009. Transport and release of chemicals from plastics to the environment and to wildlife. *Phil. Trans. R. Soc. B* 364 (1526), 2027–2045.
- Thiel, M., Gutow, L., 2005. The ecology of rafting in the marine environment. II. The rafting organisms and community. *Oceanogr. Mar. Biol. Annu. Rev.* 43, 279–418.
- Tsikhon-Lukanina, E.A., Reznichenko, O.G., Lukasheva, T.A., 2001. Feeding and spawning of the barnacle *Lepas anatifera* (Cirripedia, Lepadidae) on floating substrates in the open northwestern Pacific Ocean. *Zoologicheskii Zhurnal* 80, 650–655.
- Walkinshaw, C., Lindeque, P.K., Thompson, R., Tolhurst, T., Cole, M., 2020. Microplastics and seafood: lower trophic organisms at highest risk of contamination. *Ecotoxicol. Environ. Saf.* 190, 110066. <https://doi.org/10.1016/j.ecoenv.2019.110066>.
- Wang, S., Chen, H., Zhou, X., Tian, Y., Lin, C., Wang, W., Kaiwen, Z., Yuanbiao, Z., Lin, H., 2020. Microplastic abundance, distribution and composition in the mid-west Pacific Ocean. *Environ. Pollut.* 264, 114125. <https://doi.org/10.1016/j.envpol.2020.114125>.
- Woods, M.N., Stack, M.E., Fields, D.M., Shaw, S.D., Matrai, P.A., 2018. Microplastic fiber uptake, ingestion, and egestion rates in the blue mussel (*Mytilus edulis*). *Mar. Poll. Bull.* 137, 638–645.
- Xu, X.Y., Wong, C.Y., Tam, N.F.Y., Liu, H.M., Cheung, S.G., 2020. Barnacles as potential bioindicator of microplastic pollution in Hong Kong. *Mar. Pollut. Bull.* 154, 111081. <https://doi.org/10.1016/j.marpolbul.2020.111081>.
- Ye, S., Andrady, A.L., 1991. Fouling of floating plastic debris under Biscayne Bay exposure conditions. *Mar. Pollut. Bull.* 22, 608–613. [https://doi.org/10.1016/0025-326X\(91\)90249-R](https://doi.org/10.1016/0025-326X(91)90249-R).
- Zhang, T., Song, K., Meng, L., Tang, R., Song, T., Huang, W., Feng, Z., 2022. Distribution and characteristics of microplastics in barnacles and wild bivalves on the coast of the Yellow Sea, China. *Front. Mar. Sci.* 8, 1988.
- Zhao, S., Zhu, L., Li, D., 2015. Microplastic in three urban estuaries, China. *Environ. Pollut.* 206, 597–604. <https://doi.org/10.1016/j.envpol.2015.08.027>.